- (2) Andrew, V. E. Noble Gas Sampling System; National Technical Information Service: Springfield, VA, 1977; EMSL-LV-539-7.
- (3) Johns, F. B.; Hahm, P. B.; Thome', D. J.; Bretthauer, E. W. Radiochemical Analytical Procedures for Analysis of Environmental Samples; National Technical Information Service, U.S. Department of Commerce: Springfield, VA, 1979; EMSL-LV-0539-17.
- (4) Black, S. C.; Smith, A. E.; Costa, C. F. Offsite Monitoring for the Mighty Oak Nuclear Test; Environmental Monitoring Systems Laboratory, EPA: Las Vegas, NV, 1986; EPA 600/4-86-030.
- (5) Beiriger, J. M.; Failor, R. A.; Marsh, K. V.; Shaw, G. E. Proc. Annu. Conf. Bioassay, Anal. Environ. Radiochem. 1987, No. 33.
- (6) Meek, M. E.; Rider, B. F. Compilation of Fission Product Yields; General Electric Co., Vallecitors Nuclear Center: Pleasanton, CA, 1972; NEDO-12154.

- (7) Crust of the Earth; Poldervaart, A., Ed.; Geological Society of America: New York, 1955; pp 121-123.
- (8) Chitwood, R. B. Noble Gases; Stanley, R. E., Moghissi, A. A., Eds.; National Technical Information Service, U.S. Department of Commerce: Springfield, VA, 1973; CONF-730915, pp 69-80.
- (9) Wahlen, M.; Kunz, C. O.; Matuszek, J. M.; Mahoney, W. E.; Thompson, R. C. Science (Washington, D.C.) 1980, 207, 639-640.
- (10) International Commission on Radiological Protection Limits for Intakes of Radionuclides by Workers; Pergamon: New York, 1979; Publication 30, Suppl. Part 2, pp 374.

Received for review March 20, 1987. Revised manuscript received November 19, 1987. Accepted December 18, 1987. Mention of trade names or commercial products does not constitute endorsement or recommendation for use.

# Rate Constant for the Reaction of $NO_2$ with Sulfur(IV) over the pH Range 5.3–13

## Carol L. Clifton, Nisan Altstein, and Robert E. Hule\*

Chemical Kinetics Division, National Bureau of Standards, Gaithersburg, Maryland 20899

■ Rate constants have been determined for the reactions of NO<sub>2</sub> with SO<sub>3</sub><sup>2-</sup> and HSO<sub>3</sub><sup>-</sup> in aqueous solutions. The rate constant increases from about  $1.2 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$  near pH 5 to  $2.9 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$  at pH 13. The reaction appears to involve the formation of an intermediate complex that may undergo subsequent reaction with NO<sub>2</sub> to yield the ultimate products or may react with other substrates present. The formation of a long-lived intermediate would have implications on the chemistry of flue gas scrubbers and on luminol-based NO<sub>2</sub> detectors.

### Introduction

Air pollution by both sulfur dioxide and nitrogen oxides has been recognized as major problems almost since the start of the industrial revolution. These pollutants have been in the spotlight more recently as precursors to the strong mineral acids  $HNO_3$  and  $H_2SO_4$  in precipitation. Yet, an effective, reliable, and inexpensive technique for removing these gases from emission sources, mainly coal-fired power plants, remains an elusive goal (1). Although a considerable amount of work has been done on the development of water-based flue gas scrubbers, there are major gaps in our understanding of the basic chemistry involved in their operation, particularly the free radical chemistry. Of particular importance are the reactions of the free radical  $NO_2$ , which may be formed in the gas phase through normal combustion reactions or deliberately formed in the flue gas to aid in the removal of  $NO_x$ . In a combined  $SO_x/NO_x$  scrubber, the  $NO_2$  radical may react with dissolved S(IV) [SO<sub>2</sub>(aq), HSO<sub>3</sub><sup>-</sup>, or SO<sub>3</sub><sup>2-</sup>], effecting the removal efficiency of both.

An improved understanding of the underlying chemistry of the atmospheric conversion of  $SO_2$  to  $SO_4^{2-}$  is also needed in order to allow the development of reliable atmospheric models. In atmospheric droplets, the reaction of  $NO_2$  with  $HSO_3^-$  may serve as an additional pathway for acidification of the droplet. The reaction of  $NO_2$  with  $SO_3^{2-}$  may also play an important role in the chemistry of the luminol-based chemiluminescence detector for  $NO_2$ . The addition of  $Na_2SO_3$  has been found not only to reduce the interference from  $O_3$  but also to increase greatly the sensitivity of that detector (2, 3).

There has been some work on the effect of  $NO_2$  on  $HSO_3^{-}/SO_3^{2-}$  oxidation (4-8) and the effect of  $SO_3^{2-}$  autoxidation on the absorption of  $NO_2$  into an aqueous solution (9). Of particular interest is the observation that while  $NO_2$  will oxidize  $SO_3^{2-}$ , it does not initiate  $SO_3^{2-}$  autoxidation (8) (which is a free radical chain reaction) (10). Also, it is interesting to note that NO serves as an inhibitor of  $SO_3^{2-}$  autoxidation (6) and that the efficiency of  $NO_2$  absorption into  $SO_3^{2-}$  solutions decreased if autoxidation is permitted (9).

Recently, we investigated the reactions of NO<sub>2</sub>, generated by pulse radiolysis in aqueous solutions, with several organic and inorganic reactants (11). The pulse radiolysis apparatus employed in that study could operate only in a single-shot mode. This makes it difficult to follow very small absorbance changes. Since the optical absorption due to  $NO_2$  is very weak, its reactions were followed by monitoring a product radical from the other reactant. For the reaction with  $SO_3^{2-}$ , however, this was not possible, so the rate constant had to be measured relative to the reaction of NO2 with phenol. The results were not completely satisfactory, and only an approximate number was reported and only at pH 12.1. In this work, we have employed a pulse radiolysis apparatus with signal averaging, which has allowed us to monitor the decay of NO<sub>2</sub> directly and to measure rate constants for the reaction of  $NO_2$  with  $SO_3^{2-}$  and  $HSO_3^{-}$  over the pH range 5.3–13.

### Experimental Section

The pulse radiolysis apparatus, capable of repeated pulses under computer control, will be described in detail elsewhere (12). It employs a Van de Graaff accelerator which, in these experiments, delivers 0.5- $\mu$ s pulses of 2.8 MeV electrons to the reaction cell. The temporal history of a reactant or product is monitored by absorption spectroscopy. A decay curve from each individual pulse is recorded by a transient analyzer and transferred to the



**Figure 1.** Decay of absorbance due to NO<sub>2</sub> at 400 nm. Solution containing 10 mM NaNO<sub>2</sub> and 1.5 mM Na<sub>2</sub>SO<sub>3</sub> at pH 11.75. Average of 50 pulses. Insert: log Abs vs t.

computer. If the dose delivered by the Van de Graaff accelerator is within a preset range, the data are added to memory. This procedure is repeated enough times to produce a signal sufficient for analysis. The data are then analyzed by a linear least-squares program.

Reagent-grade chemicals were used as supplied; water was purified by an ion-exchange system. Separate solutions of NaNO<sub>2</sub> and of several concentrations of Na<sub>2</sub>SO<sub>3</sub> were prepared. The solutions were saturated with N<sub>2</sub>O and mixed just upstream of the 2 cm long irradiation cell, making use of a peristaltic pump. This procedure prevented possible complications due to the reaction of  $HNO_2$ with  $HSO_3^-$ .

The irradiation of aqueous solutions at pH 3–13 produces the hydroxyl radical and the hydrated electron in similar concentrations, along with a small (<10%) yield of hydrogen atoms. In N<sub>2</sub>O saturated solutions, the hydrated electron is converted to OH:

$$H_2O \rightarrow \rightarrow e_{ac}, H, OH$$
 (1)

$$e_{aq}^{-} + N_2 O + H_2 O \rightarrow N_2 + OH + OH^{-}$$
 (2)

The OH radical then reacts with NO<sub>2</sub><sup>-</sup> to produce NO<sub>2</sub> ( $k \sim 1 \times 10^{10} \text{ M}^{-1} \text{ s}^{-1}$ ) (13):

$$OH + NO_2^- \rightarrow OH^- + NO_2 \tag{3}$$

Under our experimental conditions, about 5  $\mu$ M NO<sub>2</sub> typically is produced. With no added solute, NO<sub>2</sub> equilibrates with its dimer N<sub>2</sub>O<sub>4</sub> ( $k = 4.5 \times 10^8$  M<sup>-1</sup> s<sup>-1</sup>;  $K = 1.53 \times 10^{-5}$  M) (14), which then disproportionates to NO<sub>2</sub><sup>-</sup> and NO<sub>3</sub><sup>-</sup> (15).

# Results and Discussion

The pulse irradiation of a  $N_2O$  saturated  $NO_2^-$  solution produced a very weak optical absorption due to NO<sub>2</sub> with maxima at 400 and 420 nm. We have not attempted to determine the absorptivity quantitatively, but our results appear compatible with the value of  $\epsilon = 200 \text{ M}^{-1} \text{ cm}^{-1}$  at 400 nm reported previously (14). To obtain an adequate signal for analysis, from 20 to 200 pulses were needed, depending upon conditions. Rate constants for the decay of the signal due to NO2 were determined upon the addition of a great excess of S(IV). A sample experimental decay curve is given in Figure 1. In all cases, good first-order behavior was observed. Replicate runs were made at each S(IV) concentration, and typically, four concentrations between  $5 \times 10^{-4}$  and  $2 \times 10^{-3}$  M were employed to derive a second-order rate constant (Figure 2). Rate measurements were carried out over the pH range 5.3-13, with the results given in Table I and plotted



**Figure 2.** Concentration dependence of the first-order rate constant for the reaction of  $NO_2$  with S(IV). Triangles, pH 5.3; squares, pH 6.7, 0.5 M phosphate buffer; diamonds, pH 11.75.



Figure 3. Dependence of the second-order rate constant for the reaction of  $NO_2$  with S(IV) on pH. Solid square, 0.5 M phosphate buffer.

Table I.	Rate	Constants	for	the	Reaction	of	$NO_2$	with
SO32-/HS	5O3-						-	

pН	$k \times 10^{-7},$ M <sup>-1</sup> s <sup>-1</sup>	conditions
5.3	1.24	1 mM nhosphate
6.7	1.54	0.5 M phosphate
6.8	1.34	$NaHSO_3 + Na_2SO_3$ , no buffer
- 8.7	1.67	1 mM borate
9.3	1.68	1 mM borate
11.8	2.14	1 mM phosphate + KOH
13	2.95	0.1 M KOH

against pH in Figure 3. Rate measurements were not attempted at lower pH to avoid complications from the reaction of  $HNO_2$  with  $HSO_3^-$ .

The pK for the deprotonation of bisulfite

$$\mathrm{HSO}_{3}^{-} \rightleftharpoons \mathrm{H}^{+} + \mathrm{SO}_{3}^{2-} \tag{4}$$

is 7.19 (16). In general, free radicals react with  $SO_3^{2-}$  more rapidly than with  $HSO_3^-$  (10, 17). This reflects the decrease in one-electron redox potential in going from  $HSO_3^$ to  $SO_3^{2-}$  and leads to an inflection in the rate constant around the pK for  $HSO_3^-$ , with a lower rate constant below pH 7.19 and a higher value above that pH. Yet, in Figure 3 there is no indication of any inflection around that point. This suggests that the change in rate constant with pH does not result simply from the increased amount of  $SO_3^{2-}$ but involves general or specific base catalysis. Indeed, although the increase in rate constant with pH appears very sharp, the total increase in going from pH 5.3 to pH 13 is only about a factor of 2.5.

Although the change in rate constant with pH is not coincident with the transition from  $HSO_3^{-1}$  to  $SO_3^{2-}$ , there is certainly an increase in rate constant with pH. To

investigate the possible importance of a general base catalyzed path, we determined the rate constant in the presence of 0.5 M phosphate buffer. The rate constant increased, but only slightly. This suggests that the increase in rate constant is associated primarily with the increase in OH<sup>-</sup> concentration. Since the increase in rate constant with pH is so small, we have not attempted to derive an OH<sup>-</sup>-dependent rate expression.

In our earlier work on the reaction of  $NO_2$  with  $SO_3^{2-}$ , we concluded that the reaction did not take place by a simple electron transfer. We arrived at this conclusion by comparing the rate constants for reactions of  $NO_2$  with comparable reactions of  $ClO_2$ . In all other cases, where electron transfer was confirmed by observing the product radicals, the reaction of  $ClO_2$  with a substrate was faster than the reaction of  $NO_2$ . The reaction of  $NO_2$  with  $SO_3^{2-}$ , however, was faster than for  $ClO_2$ . The present results support that conclusion, since if the reaction were a simple electron transfer, the change in rate constant with pH should reflect the relative amounts of  $SO_3^{2-}$  and  $HSO_3^{--}$ .

There is also other information in the literature that, in conjunction with the present observations, sheds further light on the nature of this reaction. First is the observation, mentioned above, that whereas  $NO_2$  can oxidize  $HSO_3^-$  or  $SO_3^{2^-}$  directly, it does not appear to initiate their autoxidation. Second, the products and stoichiometry suggest that the overall reaction is mainly (7)

$$2NO_2 + HSO_3^- + H_2O \rightarrow 2NO_2^- + 3H^+ + SO_4^{2-}$$
 (5)

with the yield of NO<sub>2</sub><sup>-</sup> relative to HSO<sub>3</sub><sup>-</sup> consumed 1.5 ± 0.4 and the yield of H<sup>+</sup> 2.5 ± 0.4. There appeared to be essentially no NO or NO<sub>3</sub><sup>-</sup> produced. Third, the rate constant (d[H<sup>+</sup>]/dt) determined by bubbling NO<sub>2</sub> through a HSO<sub>3</sub><sup>-</sup> solution (>2 × 10<sup>6</sup> M<sup>-1</sup> s<sup>-1</sup> at pH 6.4 and 1.4 × 10<sup>5</sup> M<sup>-1</sup> s<sup>-1</sup> pH 5) (18) was much lower than the rate constant determined here. That work also showed that the reaction was not second order in NO<sub>2</sub>, as the stoichiometry would suggest. Finally, it was observed that there was a small induction period in the generation of H<sup>+</sup> and that the generation of H<sup>+</sup> would continue for several minutes after NO<sub>2</sub> had been purged from the solution.

In our earlier work, we suggested that the reaction was a simple oxygen-atom transfer:

$$NO_2 + HSO_3^- \rightarrow NO + SO_4^{2-} + H^+$$
(6)

The observations summarized above suggest that this is not likely to be the case. Rather, our observations and these other results suggest that the initial step involves the formation of an addition complex, which can undergo subsequent reaction with further  $NO_2$ :

$$NO_{2} + SO_{3}^{2^{-}} \longrightarrow \bigcup_{0}^{1^{-}} O^{-} S \longrightarrow O^{-} (7)$$

$$NO_{2} + \bigcup_{0}^{1^{-}} O^{-} S \longrightarrow O^{-} \bigcup_{0}^{1^{-}} O^{-} S \longrightarrow O^{-} \bigcup_{0}^{1^{-}} O^{-} \bigcup_{0}^{1^{-}}$$

This second intermediate, which is potentially long lived, can then decompose, possibly in a base-catalyzed reaction, to  $NO_2^-$  and  $SO_4^{2-}$ :

In our experiments, the low initial NO<sub>2</sub> concentration  $(<10^{-5} \text{ M})$  will probably ensure that reaction 7 predominates. The occurrence of reaction 8 cannot be ruled out completely, however, but the very good first- and second-order plots (Figures 1 and 2) suggest that it is not important. Therefore, the rate constant we measure should

588 Environ. Sci. Technol., Vol. 22, No. 5, 1988

correspond to the formation of the first complex. Indeed, there does seem to be a very small residual absorption at long time (Figure 1) that could be due to this complex. The production of acid upon bubbling  $NO_2$  through a  $HSO_3^-$  solution, however, requires reactions 8–10 and would be expected to be much slower, as observed (18).



Relatively slow secondary steps also would be expected to lead to an induction period for the production of acid and the continued production of acid after the  $NO_2$  was removed.

There is certainly considerable precedence for an addition reaction in the interactions of sulfite and bisulfite with nitrogen oxides and their anions. Nitrous acid reacts with bisulfate, first to form nitrososulfonic acid (19), which can then react further with bisulfate to form hydroxylaminedisulfonate (20):

$$HNO_2 + HSO_3^- \rightarrow ONSO_3^- + H_2O \qquad (11)$$

$$ONSO_3^- + HSO_3^- \rightarrow HON(SO_3)_2^{2-} + H_2O \quad (12)$$

This product may hydrolyze or react with additional HSO<sub>3</sub><sup>-</sup> to form aminetrisulfonate:

$$HON(SO_3)_2^{2-} + HSO_3^{-} \to N(SO_3)_3^{3-} + H_2O$$
 (13)

In these reactions, three sulfites substitute onto the nitrogen, with the elimination of water. The reaction of nitric oxide with sulfite (21, 22), however, appears to be more similar to what we are proposing here for nitrogen dioxide:

$$NO + SO_3^{2-} \rightarrow ONSO_3^{2-} \tag{14}$$

$$NO + ONSO_3^{2-} \rightarrow ON(NO)SO_3^{-}$$
(15)

$$-ON(NO)SO_3^- \rightarrow N_2O + SO_4^{2-}$$
(16)

In an atmospheric droplet, the fate of the initial intermediate formed from the reaction of NO<sub>2</sub> with SO<sub>3</sub><sup>2-</sup> might not be simply to react with another NO<sub>2</sub>, since that radical will be at such a low concentration in the droplet. The compound might serve as a one-electron oxidant, leading to the formation of nitrososulfuric acid. In a flue gas scrubber, particularly when the gas-phase NO is converted to NO<sub>2</sub>, the intermediate is much more likely to react with additional NO<sub>2</sub> due to its much higher concentration. Confirmation of the production of this intermediate and information on its chemical behavior are needed to fully assess its role in these systems.

The formation of a reactive intermediate in the reaction of NO<sub>2</sub> with SO<sub>3</sub><sup>2-</sup> might also help explain the role of added Na<sub>2</sub>SO<sub>3</sub> on the operation of a luminol-based NO<sub>2</sub> detector. The reaction of SO<sub>3</sub><sup>2-</sup> with O<sub>3</sub> is a simple oxygen-atom transfer (23), leading to unreactive products:

$$O_3 + SO_3^{2-} \to O_2 + SO_4^{2-}$$
 (17)

This explains the ability of  $SO_3^{2-}$  to reduce the interference due to  $O_3$ . When  $NO_2$  is dissolved into a luminol solution, its reaction with luminol is in competition with the removal

of  $NO_2$  due to disproportionation. If a large concentration of  $SO_3^{2-}$  is present, the NO<sub>2</sub> will react to form an intermediate. This intermediate, then, can proceed to react with the luminol, leading to chemiluminescence. An investigation into the nature of this intermediate would be useful in predicting possible complications in the luminol-based  $NO_2$  detector.

## Acknowledgments

We thank Dr. G. W. Harris of the Max Plank Institute for Chemistry, Mainz, FRG, for alerting us to the use of sulfite in the luminol-based  $NO_2$  detector.

## **Registry No.** NO<sub>2</sub>, 10102-44-0; NO<sub>x</sub>, 11104-93-1.

# Literature Cited

- (1) Chang, S. G. ACS Symp. Ser. 1986, No. 319, 159
- Maeda, Y.; Aokl, K.; Munemorl, M. Anal. Chem. 1980, 52, (2)307.
- Wendel, G. J.; Stedman, D. H.; Cantrell, C. A. Anal. Chem. (3)1983, 55, 937.
- (4) Nash, T. Atmos. Environ. 1979, 13, 1149.
- (5) Rosenberg, H. S.; Grotta, H. M. Environ. Sci. Technol. 1980, 14, 470.
- (6) Cofer, W. R.; Schryer, D. R.; Rogowski, R. S. Atmos. Environ. 1981, 15, 1281.
- (7) Lee, Y. N.; Schwartz, S. E. In Precipitation Scavenging, Dry Deposition, and Resuspension; Pruppacher, H. R., Semonin, R. G., Slinn, W. G. N., Eds.; Elsevier: New York, 1983; p 453.
- (8) Ellison, T. K.; Eckert, C. A. J. Phys. Chem. 1984, 88, 2335.
- (9) Takeuchi, H.; Takahashi, K.; Kizawa, N. Ind. Eng. Process Des. Dev. 1977, 16, 486.

- (10) Neta, P.; Huie, R. E. EHP, Environ. Health Perspect. 1985, 64, 209.
- (11) Huie, R. E.; Neta, P. J. Phys. Chem. 1986, 90, 1193.
- (12) Huie, R. E.; Clifton, C. L.; Altstein, N., submitted for publication in Radiat. Phys. Chem.
- (13) Farhataziz; Ross, A. B. Natl. Stand. Ref. Data. Ser. (U.S. Natl. Bur. Stand.) 1977, NSRDS-NBS 59.
- (14) Graetzel, M.; Henglein, A.; Lilie, J.; Beck, G. Ber. Bunsen-Ges. Phys. Chem. 1969, 73, 646.
- (15) Schwartz, S. E.; White, W. H. In Trace Atmospheric Constituents: Properties, Transformations, and Fates; Schwartz, S. E., Ed.; Wiley: New York, 1983; p 1. (16) Goldberg, R. N.; Parker, V. B. J. Res. Natl. Bur. Stand.
- (U.S.) 1985, 90, 341.
- (17) Huie, R. E.; Neta, P. Atmos. Environ. 1987, 21, 1743.
- (18) Lee, Y. N.; Schwartz, S. E., Brookhaven National Laboratory, ref 7 and private communication.
- (19) Oblath, S. B.; Markowitz, S. S.; Novakov, T.; Chang, S. G. J. Phys. Chem. 1982, 86, 4853.
- (20) Oblath, S. B.; Markowitz, S. S.; Novakov, T.; Chang, S. G. J. Phys. Chem. 1981, 85, 1017.
- (21) Nunes, T. L.; Powell, R. E. Inorg. Chem. 1970, 9, 1916.
- (22) Littlejohn, D.; Hu, K. Y.; Chang, S. G. Inorg. Chem. 1986, 25. 3131.
- (23) Hoffman, M. R. Atmos. Environ. 1986, 20, 1145.

Received for review June 25, 1987. Accepted December 14, 1987. Although the research described in this paper has been funded in part by the U.S. Environmental Protection Agency through Interagency Agreement DW13931327-01-0, it has not been subjected to the Agency's required peer and policy review and therefore does not necessarily reflect the views of the Agency, and no official endorsement should be inferred.